

A Simulation of air quality in association with haze events occurred in East Asia during 10-15 January 2013

Soon-Ung Park and Jeong Hoon Cho

Abstract—A prolonged heavy haze event that has caused for the Environmental Protection Bureau (EPB) in Beijing to take emergency measures for the protection of the public health and the reduction of air pollution damages in China has been analyzed with the use of the Aerosol modeling System (AMS) to identify causes of this event. It is found that the heavy haze event is associated with high aerosols and water droplets concentrations. These high aerosol concentrations are mainly composed of anthropogenic aerosols, especially secondary inorganic aerosols formed by gas-to-particle conversion of gaseous pollutants in the eastern part of China whereas those in the northeastern parts of China are composed of the mixture of the anthropogenic aerosols and the Asian dust aerosol originated from the dust source regions of northern China and Mongolia. These high aerosol concentrations are found to be subsequently transported to the downwind regions of the Korean Peninsula and Japan causing a prolonged haze event there. It is also found that the Asian dust aerosol from northern China and Mongolia and the anthropogenic aerosols produced by chemical reactions of pollutants in eastern China can cause significantly adverse environmental impacts in the whole Asian region by increased atmospheric aerosol loadings that may cause respiration diseases and visibility reduction and by excess deposition of aerosols causing adverse impacts on terrestrial and marine eco-systems. High concentrations of gaseous pollutants of SO₂ and NO₂ can enhance poor air quality over the high emission regions, however, the O₃ concentration tends to decrease over the high NO_x emission region but to increase over the low NO_x emission region thereby sometimes resulting in high O₃ concentration in the downwind region of oceans.

Keywords—Aerosol loading and deposition, Aerosol Modeling System (AMS), Anthropogenic aerosol, Asian Dust Aerosol Model 2 (ADAM2), CMAQ, Pollutants emissions in Asia.

I. INTRODUCTION

AIR quality is defined as a measure of the condition of air relative to the requirements of one or more biotic species and/or human need or purpose [1]. Poor air quality in East Asia has become a major environmental problem in recent years due to rapid economical growths in most of Asian countries. The air quality is largely determined by concentrations of gaseous

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pollutants (SO₂, NO_x, CO, O₃) and atmospheric aerosols [2]-[8].

Atmospheric aerosols can affect the quality of our lives significantly because of its potential impacts on human health and the environment. The submicrometer size of aerosols can be inhaled and thus may pose certain health hazards [9]-[16]. Because aerosols also scatter light, they strongly influence the radiative budget of the Earth-atmosphere system; they also reduce visibility and diminish the aesthetic scenery [17]—[26]. Visibility reduction is usually caused by weather phenomena such as precipitation, fog, mist, haze and dust that are associated with hydrometeor and lithometeor [27].

East Asia is a major source of both natural aerosol (Asian dust) and anthropogenic aerosols over the Northern Hemisphere. Asian dust that is a typical example of mineral aerosol occurs in northern China and Mongolia more frequently during the spring season [28]-[33] and has its increasing occurrence trend due to desertification in the source region. Anthropogenic aerosols that are mainly originated from emitted pollutants have also an increasing trend due to the rapid economic expansion in many Asian countries [34]-[37]. Tropospheric aerosols in this region are the complex mixture of various aerosols such as Asian dust and anthropogenic aerosols. Consequently, occurrence frequencies of visibility reducing weather events caused by aerosols have an increasing trend, especially over Asia [38]-[42].

Visibility reducing weather phenomena caused by aerosols include precipitation, fog, mist, haze and dust storm [27]. Among these, mist and haze that are composed of submicrometer particles are mainly formed by gas-to-particle conversion processes in the atmosphere [43]-[45]. The recent increasing occurrence frequency of dense haze and mist events in the eastern parts of China appears to be related to this size range of secondary inorganic aerosols that are formed through the chemical reactions of gaseous pollutants such as SO₂, NO_x, NH₃ and water vapor.

In fact a wide swath of central and eastern China experienced several days of the worst air pollution the country has seen in recent memory, with the dense haze covering several provinces in China including Beijing, Hebei, Tianjing, Shandong, Henan, Jiangsu, Anhui, Jiangxi and Hubei from 11 to 16 January 2013. These events caused for the Environmental Protection Bureau (EPB) in Beijing to take emergency measures including halting

outdoor activities for primary and middle school students, suspending construction of 28 construction sites, reducing emissions by 30 % at 58 factories, and taking up to 30 % of government vehicles off the road (China Daily). Haze (Smog) blanketed Shandong and Jiangxi provinces has forced the closure of many highways and the cancellation or delay of many air flights during these events period. These dense haze events have been transported downwind regions causing high aerosol concentrations of more than $200 \mu\text{g}/\text{m}^3$ over Korea and $50 \mu\text{g}/\text{m}^3$ in parts of Japan.

Recently the Aerosol Modeling System (AMS) that is composed of the Asian Dust Aerosol Model 2 (ADAM2) for the Asian dust aerosol modeling and the Community Multiscale Air Quality (CMAQ) model for the anthropogenic aerosol modeling has been developed [37]. AMS could successfully simulate a prolonged dense haze event occurred during the period of 19-22 May 2010 [46] in East Asia.

The Purpose of this study is examine the air quality in association with dense haze events occurred for the period from 10 to 15 January 2013 in East Asia with the use of Aerosol Modeling System (AMS) and to investigate the effects of various aerosol species and pollutants on these hazardous environmental events.

II. MODEL DESCRIPTIONS

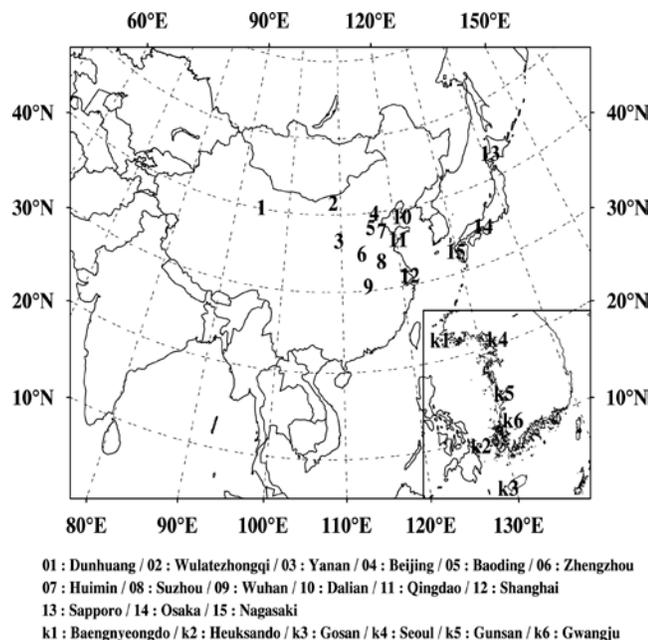


Fig. 1 The model domain with the indication of sites that will be described in the text. The enhanced map of South Korea with the indication of monitoring sites is shown in the right lower corner of the domain.

A. Meteorological model

The meteorological model used in this study is the fifth generation mesoscale model of non-hydrostatic version (MM5; Pennsylvania State University / National Center for

Atmospheric Research) defined in the x , y and σ coordinate [47][48]. The model domain (Fig. 1) has the horizontal resolution of $27 \text{ km} \times 27 \text{ km}$ with 30 vertical layers in the Asian region.

The NCEP FNL operational global analysis data on $1.0 \text{ degree} \times 1.0 \text{ degree}$ grids are used for the initial and lateral boundary conditions for the model.

B. Aerosol Modeling System (AMS)

The Aerosol Modeling System (AMS) is consisted of the Asian Dust Aerosol Model 2 (ADAM2) [49] and the Community Multiscale Air Quality (CMAQ) modeling system (<http://www.cmaq-model.org>) with emission data of pollutants (SO_2 , NO_x , VOC, CO, NH_3 , BC, OC and PM_{10}) in the model domain.

C. ADAM2

The ADAM2 model is an Eulerian dust transport model that includes the specifications of the dust source regions delineated by the statistical analysis of the World Meteorological Organization (WMO) 3 hourly reporting dust data and statistically derived dust emission conditions in Sand, Gobi, Loess and Mixed surface soil in the model domain (Fig. 1). The model uses the suspended particle-size distribution parameterized by the several log-normal distributions in the source regions, based on the parent soil particle-size distributions with the used of the concept of the minimally and fully dispersed particle-size distribution [32][50]-[52]. It has 11-size of bins with near the same logarithm interval for particles of $0.15\text{-}35 \mu\text{m}$ in radius [31][32]. The model has a temporally varying emission reduction factors derived statistically using a normalized difference vegetation index (NDVI) in the different surface soil types in the Asian dust source region. The detailed description is given in [49].

D. CMAQ model

The Environmental Protection Agency (EPA) Community Multiscale Air Quality (CMAQ) modeling system (<http://www.cmaq-model.org>) is a three-dimensional eulerian atmospheric chemistry and transport modeling system that simulates airborne pollutants, ozone concentration, particulate matters, visibility, and acidic and nutrient pollutant species throughout the troposphere [53].

The aerosol component of the CMAQ model has the particle size distribution as the superposition of three lognormal subdistributions, called modes. Fine particles with diameters less than $2.5 \mu\text{m}$ ($\text{PM}_{2.5}$) are represented by two subdistributions called the Aitken and accumulation modes. The Aitken mode includes particles with diameters up to approximately $0.1 \mu\text{m}$ for the mass distribution and the accumulation mode covers the mass distribution in the range from 0.1 to $2.5 \mu\text{m}$. The coarse mode covers the mass distribution in the range from 2.5 to $10 \mu\text{m}$. The model includes the processes of coagulation, particle growth by the addition of mass and new particle formation [54].

E. Emission data

Air pollutant emission data (Fig. 2) in the year 2010 are

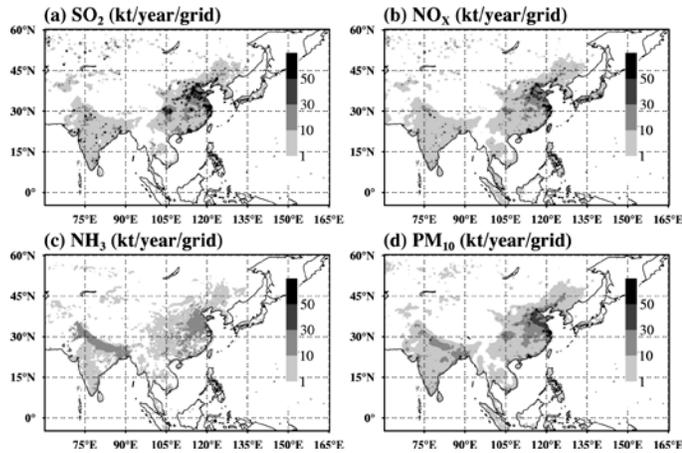


Fig. 2 Horizontal distributions of emission rate (kt/yr/grid) of (a) SO₂, (b) NO_x, (c) NH₃ and (d) PM₁₀ in the year 2010 over Asian region (grid: 0.5° × 0.5°).

obtained from the International Institute for Applied Systems Analysis (IIASA; ftp://www.iiasa.ac.at/outgoing/may/KRF) in a grid of 0.5° long. × 0.5° lat. in the global domain. The emission data include SO₂, NO_x, NH₃, CO, VOC, PM₁₀, BC and OC. The estimated total Asian (China's) anthropogenic emissions in the year 2010 are 48.4 Tg (30.0 Tg) SO₂, 36.7 Tg (14.0 Tg) NO_x, 43.5 Tg (22.4 Tg) VOC, 279.0 Tg (144.5 Tg) CO, and 36.5 Tg (23.0 Tg) PM₁₀, suggesting more than 50 % of the total anthropogenic emissions being contributed by China in the Asian domain.

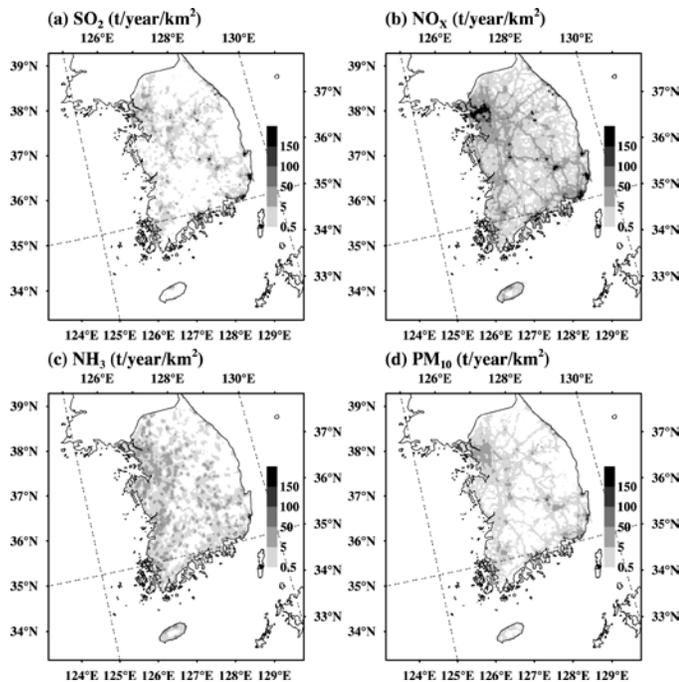


Fig. 3 Horizontal distributions of emission rate (t/yr/grid) of (a) SO₂, (b) NO_x, (c) NH₃, and (d) PM₁₀ (1 grid: 1 km × 1 km).

Air pollutant emissions in South Korea (Fig. 3) in the year 2007 are obtained from the Clean Air Policy Supporting System (CAPSS, Korean Ministry of Environment) in a grid of 1 km × 1

km. The estimated total anthropogenic emission in the year 2007 are 402,525 t SO_x, 1,187,923 t NO_x, 874,699 t VOC, 808,862 t CO, and 98,143 t PM₁₀. These emission data over South Korea (Fig. 3) are used for the model simulation rather than using the emission inventory data of IIASA.

III. SYNOPTIC SITUATIONS OF THE CHOSEN HAZE EVENTS

Fig. 4 shows the surface weather analysis map obtained from Korea Meteorological Administration (KMA, 2013) during the

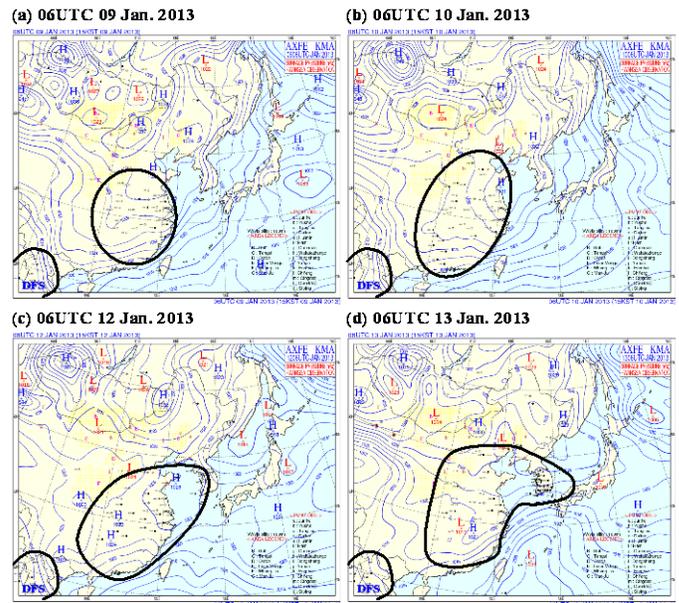


Fig. 4 Surface weather analysis maps at (a) 06 UTC 9, (b) 06 UTC 10, (c) 06 UTC 12 and (d) 06 UTC 13 January 2013. The haze is indicated by "∞". The circle shows the haze occurrence regions.

haze event period in East Asia.

The most parts of eastern China are under the influence of a surface high pressure system centered at central northern China (110°E and 40°N) at 00 UTC 9 January 2013. As this high pressure system moves slowly southeastward, haze is reported in the wide region of eastern China at 06 UTC 9 January (Fig. 4a). Thereafter this surface high pressure system continuously moves further southeastward to locate in the coastal region of Shandong province at 06 UTC 10 January (Fig. 4b), the dense haze covers almost all over eastern China from northeast China to the South China Sea. At 06 UTC 12 January the surface pressure center located at Shandong province slightly moves northeastward over the Yellow Sea accompanied with further northeastward extension of the dense haze zone over South Korea (Fig. 4c).

Thereafter the dense haze occurrence zone is prevailed over South Korea and extended over Japan with the surface high pressure system is keeping to move northeastward over northeastern China (Fig. 4d). This trend is continued until 18 UTC 15 January when the surface low pressure system developed over the East China Sea moves northeastward to the East Sea of Korea. This prolonged dense haze event has caused for the Environmental Protection Bureau of China to take

emergency measures in several provinces in eastern China.

IV. RESULTS OF THE MODEL SIMULATION

A. Comparison of observed and simulated aerosol (PM_{10}) concentrations over South Korea

The Aerosol Modeling System (AMS) has been employed to simulate concentrations of PM_{10} and pollutants for the period from 4 to 16 January 2013 that includes the dense haze event period in East Asia in the domain given in Fig. 1.

The time series of observed PM_{10} and gas phase pollutants (SO_2 , NO_2 and O_3) concentrations at several monitoring sites over South Korea (Fig. 1) are compared with the model simulated ones to see the performance of the model.

Fig. 5 shows time series of hourly mean surface PM_{10} concentrations observed and simulated by the model at several monitoring sites over South Korea (Fig. 1). The simulated PM_{10} concentrations are composed of all kinds of aerosols including Water (water droplet that is formed through the chemical reaction and hygroscopic processes), Other (sea salt and secondary organic aerosols), Asian dust (aerosols emitted from soil erosion), SIA (secondary inorganic aerosols; SO_4^{2-} , NO_3^- , NH_4^+), BC (black carbon) and OC (organic carbon), and unspecified PM_{10} (emitted anthropogenic aerosol not include above categories).

It is worthwhile to note that the observed PM_{10} concentration in Fig. 5 does not include the water droplet aerosol since the sampled air is desiccated before measuring PM_{10} concentration.

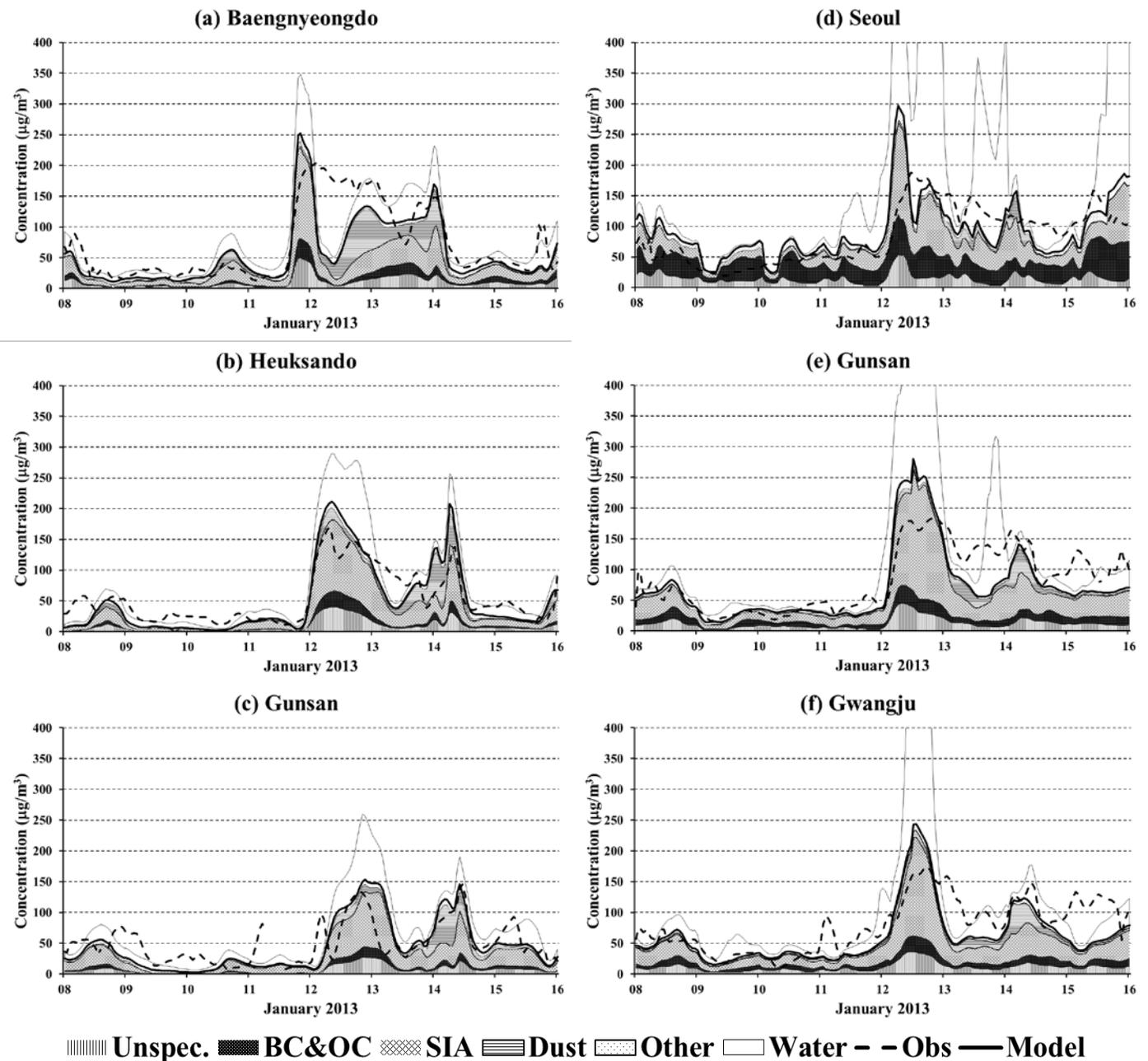


Fig. 5 Time series of hourly mean observed (dashed line) and modeled (various shadings) PM_{10} concentration ($\mu g/m^3$) at (a) Baengnyeongdo, (b) Heuksando, (c) Gosan, (d) Seoul, (e) Gunsan and (f) Gwangju in Korea for the period of 09:00 LST 8 to 09:00 LST 16 January 2013.

Therefore, the comparison between the measured and simulated PM₁₀ concentrations should be made without water droplet aerosols in the simulated PM₁₀ concentration. However, the water droplet aerosol concentration is very important to distinguish the hygroscopic aerosol from non-hygroscopic aerosol and the mist from haze. It also affects the visibility significantly [27].

The model simulates quite well observed high PM₁₀ concentration events that caused dense haze events over Korea. Two high aerosol-concentration events are simulated (Fig. 5). The first event occurs at 00:00 LST (= UTC + 9) 12 January at Baengnyeongdo (Fig. 1) with the surface maximum PM₁₀ concentration more than 230 μg/m³ at 05:00 LST (Fig. 5a). This event moves southeastward to the Gosan site in Jeju island of

Korea (Fig. 1) at 16:00 LST 12 January 2013 with the maximum PM₁₀ concentration of 150 μg/m³ at 05:00 LST 13 January (Fig. 5c). While at the inland site of Gwangju that is located about 500 km to the southeast of Baengnyeongdo (Fig. 1) the event occurs at 09:00 LST 12 January with the simulated maximum PM₁₀ concentration exceeding 200 μg/m³ at 00:00 LST 13 January (Fig. 5f). The observed and simulated high aerosol concentrations of the first event are mainly contributed by the anthropogenic aerosols (Fig. 5). Among these aerosols the secondary inorganic aerosols (SO₄²⁻, NO₃⁻, and NH₄⁺) are predominated, suggesting the importance of converted aerosols from emitted air pollutants.

The second high aerosol concentration event occurs at 16:00 LST 12 January at the Baengnyeongdo site with the maximum

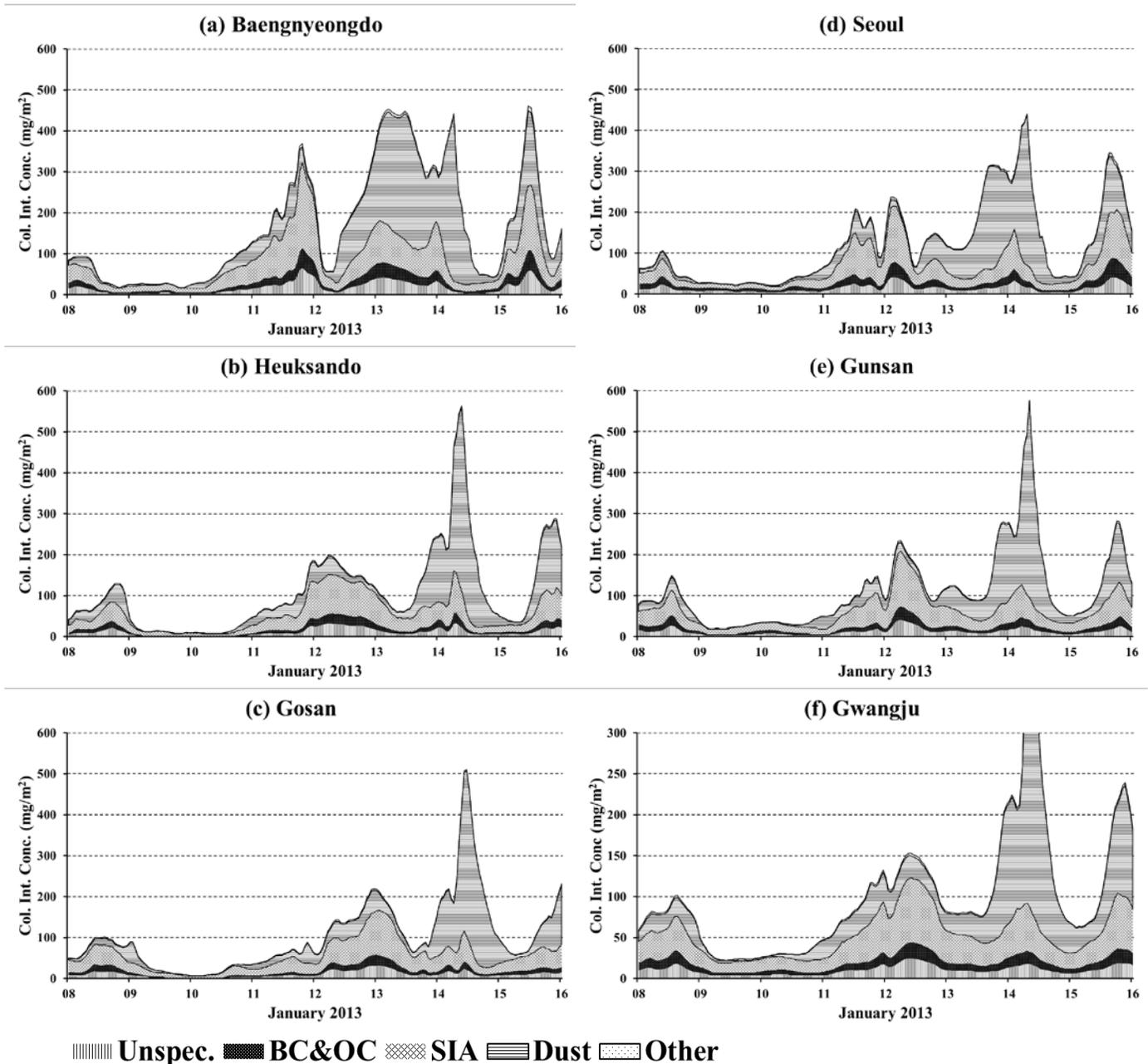


Fig. 6 The same as in Fig. 5 except for the column integrated PM₁₀ concentration (mg/m²)

surface PM_{10} concentration of more than $150 \mu\text{g}/\text{m}^3$ at 10:00 LST 13 January and ends at around 15:00 LST 14 January 2013 (Fig. 5a). This event moves slowly southeastward to Gosan (starting time at 00 LST 14 and ending time at 00 LST 15), with maximum surface PM_{10} concentration of $150 \mu\text{g}/\text{m}^3$ (Fig. 5c) and to Gwangju (starting time at 06:00 LST 14 and ending time at 16:00 LST 15 January) with the maximum surface PM_{10} concentration of about $150 \mu\text{g}/\text{m}^3$ (Fig. 5f). During the second event period hazes are reported all over South Korea (Figs. 4c and d), suggesting the haze event reported in South Korea being caused by the high aerosol concentration.

The second high aerosol-concentration event in Fig. 5 is contributed by the mixture of Asian dust aerosol and the anthropogenic aerosols that is different from the first event which is mainly caused by the anthropogenic aerosol. The difference between the first and second high aerosol-concentration events is more clearly seen in Fig. 6. The column integrated PM_{10} concentration (Fig. 6) indicates that the first high aerosol-concentration event is mainly contributed by anthropogenic aerosols, especially by secondary inorganic aerosols (SIA) originated from air pollutants, whereas the second event is largely contributed by the Asian dust aerosol originated from the Asian dust source region at all monitoring sites in Korea (Figs. 6a, b, c, d, e, and f). Note that the upper level long-range transported Asian dust aerosol is largely attributed to the haze phenomena of the second event; the column integrated concentration of the Asian dust aerosol is much greater than that of the anthropogenic aerosols for the second event (Fig. 6) whereas that of the anthropogenic aerosol is predominated for the first event (Fig. 6). The surface weather analysis map in Fig. 4c clearly indicates that the first high aerosol concentration is related with the haze reports over the western parts of the Korean peninsula.

B. Comparison of observed and simulated gas phase pollutants

Figs. 7, 8 and 9, respectively show time series of observed and simulated SO_2 , NO_2 and O_3 concentrations at several monitoring sites over South Korea (Fig. 1). The model simulates quite well the observed SO_2 , NO_2 , and O_3

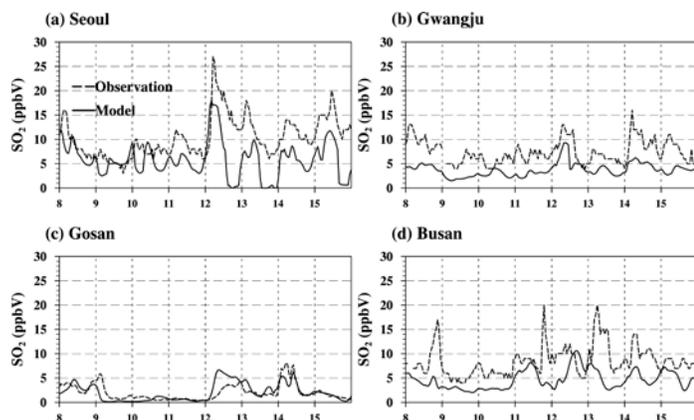


Fig. 7 Time series of observed (dashed line) and modeled (solid line) SO_2 concentration (ppb) at (a) Seoul, (b) Gwangju, (c) Gosan and (d) Busan for the period of from 8 to 16 January 2013.

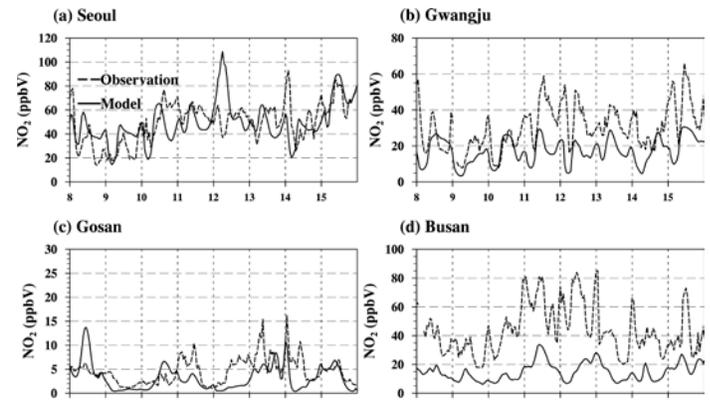


Fig. 8 The same as in Fig. 7 except for NO_2 concentration (ppb).

concentrations at Seoul (Figs. 7a, 8a and 9a), Gosan (Figs. 7c, 8c and 9c). However, the model underestimates SO_2 and NO_2 concentrations at Gwangju (Figs. 7b and 8b) and Busan (Figs. 7d and 8d). This may be due to the underestimated emission amounts of SO_2 and NO_x around these sites. Underestimation of NO_2 concentration (Figs. 8b and d) leads to the overestimation of O_3 concentration as seen in Figs. 9b and d.

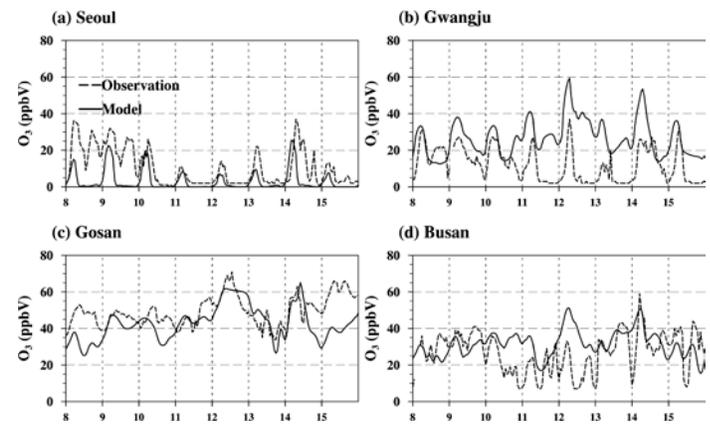


Fig. 9 The same as in Fig. 7 except for O_3 concentration (ppb).

C. Temporal variations of model simulated PM_{10} concentration in China and Japan

Fig. 10 shows time series of hourly mean surface PM_{10} concentration at three sites in the Asian dust source region (Fig. 1 and Figs. 10a-c) and three sites at the eastern border of the Asian dust source region in China (Fig. 1 and Figs. 10d-f). Asian dust occurs every day during the analysis period starting from 00 LST (= UTC + 8) 11 January with varying intensity depending on the site and time (Figs. 10a-c). The maximum hourly mean PM_{10} concentration of $1,800 \mu\text{g}/\text{m}^3$, $150 \mu\text{g}/\text{m}^3$, $200 \mu\text{g}/\text{m}^3$, respectively occurs at 00 LST 13 at the Dunhung site (Fig. 10a), 00 LST 15 at the Wulantezhongzi site (Fig. 10b) and 04 LST 11 January at the Yanan site (Fig. 10c). Most of aerosols at these sites are contributed by the Asian dust aerosol (Figs. 10a-c).

However, toward the border of the Asian dust source region, the contribution of anthropogenic aerosols becomes more important as seen in Figs. 10d-f. The high anthropogenic aerosol concentration of more than $250 \mu\text{g}/\text{m}^3$ at Beijing (Fig. 10d), and more than $400 \mu\text{g}/\text{m}^3$ at Baoding (Fig. 10e) and

Zhengzhou (Fig. 10f) throughout the analysis period together with the high water droplet aerosol concentration has caused dense haze events reported in Fig. 4. The most predominant aerosol for this haze event is the secondary inorganic aerosols that are formed by gas-to-particle conversion processes in the atmosphere, suggesting the importance of SO₂, NO_x and NH₃ emissions.

Fig. 11 shows the time series of model simulated hourly mean PM₁₀ concentration at three sites in the central eastern low land in China (Figs. 11a-c) where pollutants emissions are high (Fig. 2), and the three eastern coastal sites in China (Figs. 11d-f). More than 300 µg/m³ of PM₁₀ concentration occurs during the period from 10 to late 14 January at the site located in the

northern low land region (Fig. 11a) while at the other sites located in the low land (Figs 11b and c) the surface PM₁₀ concentration increases throughout the analysis period with the maximum PM₁₀ concentration of more than 500 µg/m³. Aerosols are mainly composed of anthropogenic aerosols including SIA, BC, OC and unspecified PM₁₀ that are originated from the pollutant emissions. The high concentrations of anthropogenic aerosols (especially the secondary inorganic aerosol) together with that of the water droplet aerosol have caused the prolonged haze event in this region of China (Fig. 4).

At the Dalian site (Fig. 11d) located in the northeastern coast of China (Fig. 1), two main heavy aerosol events occur for the periods of 06 LST 10 to 00 LST 12 January and 09 LST 12 to 09

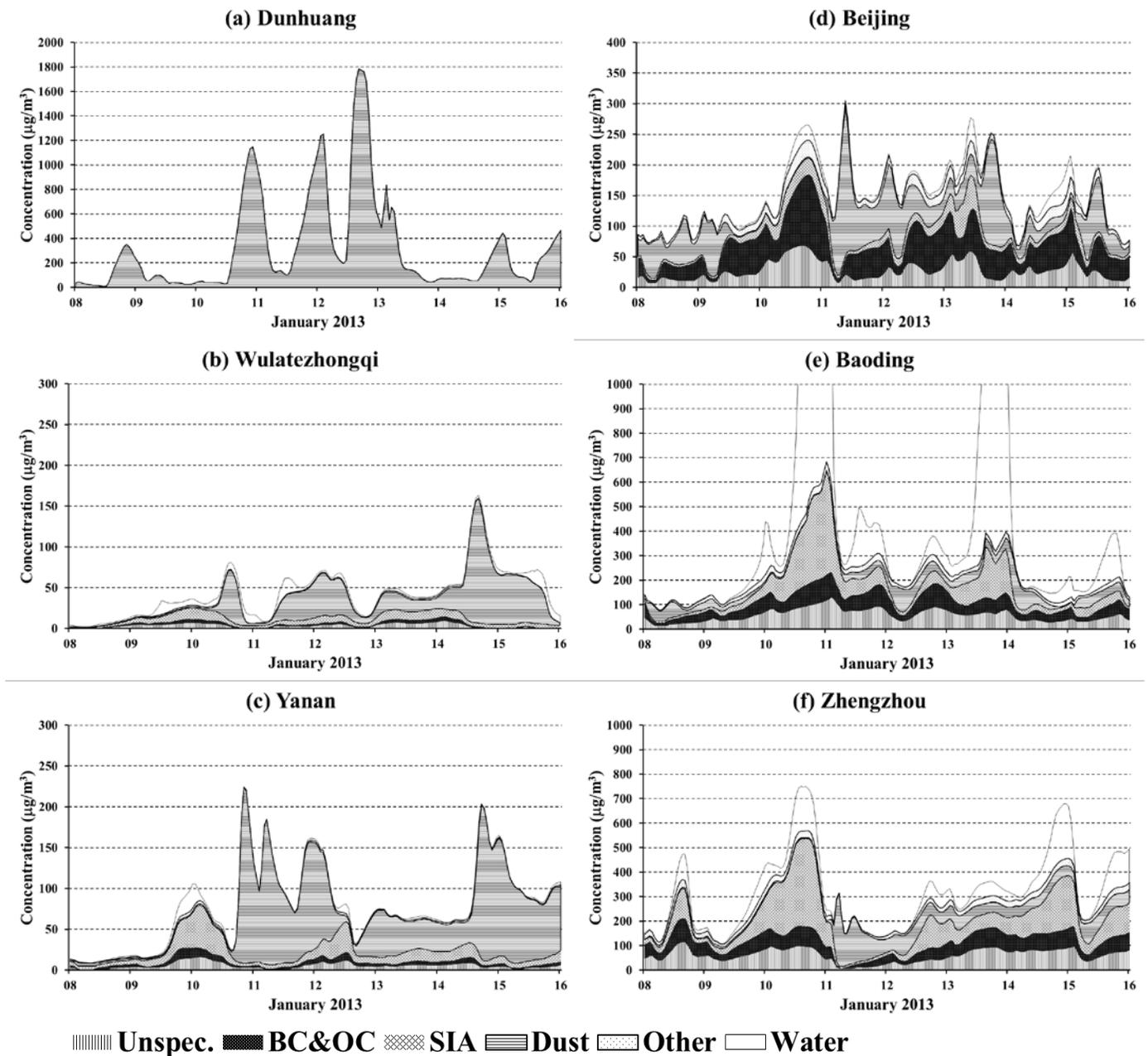


Fig. 10 The same as in Fig. 5 except for the model simulated PM₁₀ concentration (µg/m³) at (a) Dunhuang, (b) Wulatezhongqi and (c) Yanan located in the Asian dust source region, and at (d) Beijing, (e) Baoding and (f) Zhengzhou located near the eastern border of the Asian dust source region.

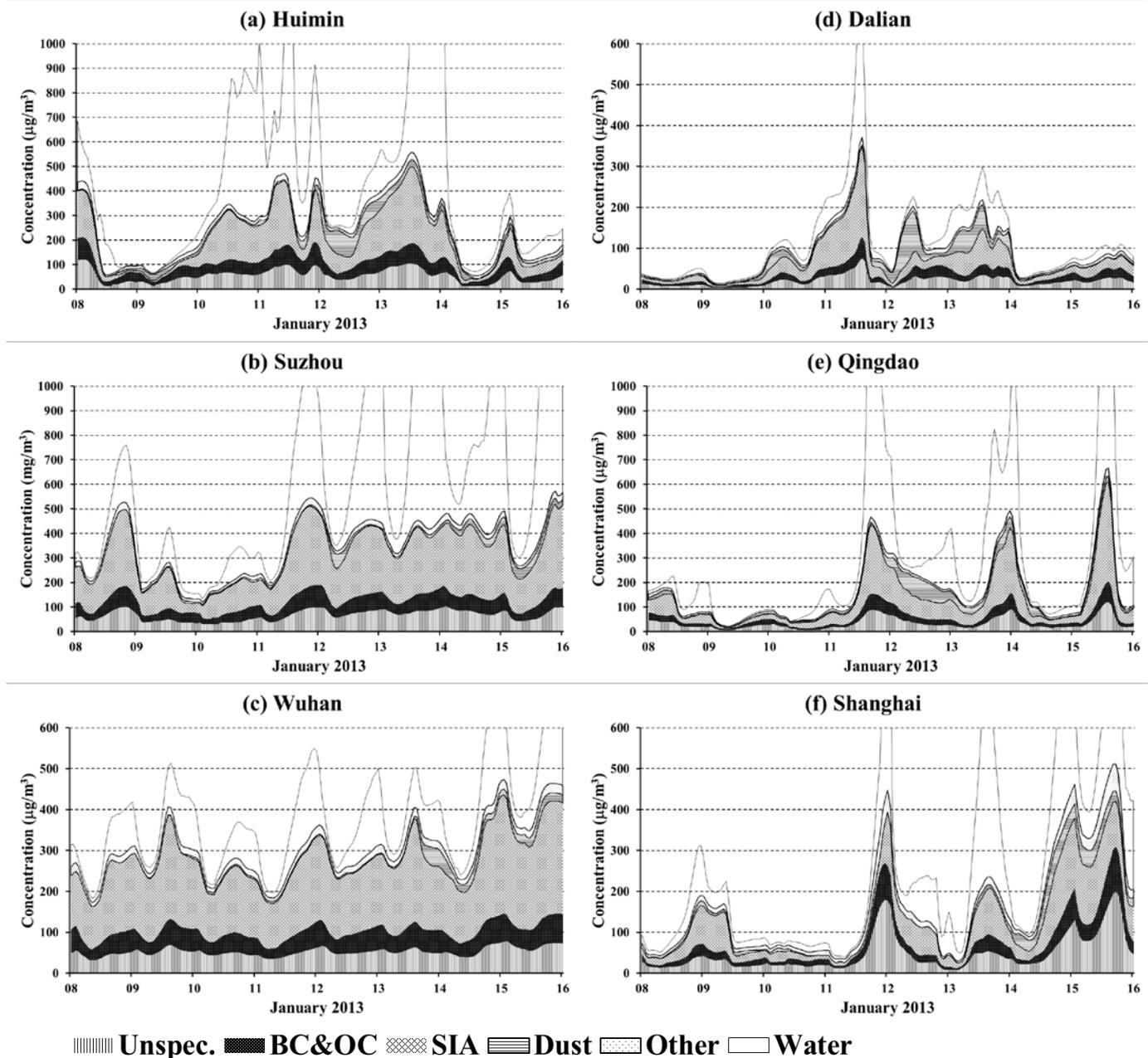


Fig. 11 The same as in Fig. 10 except at (a) Huimin, (b) Suzhou and (c) Wuhan located in the eastern low flat region, and at (d) Dalian, (e) Qingdao and (f) Shanghai located in the eastern coastal region of China.

LST 14 January. The first one is mainly contributed by anthropogenic aerosols with the maximum surface PM_{10} concentration of more than $300 \mu\text{g}/\text{m}^3$ at 23 LST 11 January, while the second one is contributed by the mixture of anthropogenic aerosols and the Asian dust aerosol originated from northern China.

At the coastal sites of Qingdao and Shanghai (Figs. 11e and f) the high aerosol concentration event has occurred before 00 LST 10 January as is the case at sites in the high pollutants emission region (Figs. 11a, b and c) but the main event starts to occur from 14 LST 11 January with increasing intensity throughout the analysis period (Figs. 11e and f). The maximum surface PM_{10} concentration is more than $600 \mu\text{g}/\text{m}^3$ at 22 LST 15 at Qingdao and $420 \mu\text{g}/\text{m}^3$ at 00 LST 16 January at Shanghai

causing dense haze events in these regions. Aerosols are mainly composed of anthropogenic aerosols for the first event occurred at 00 LST 12 and the third event at 22 LST 15 January at the Qingdao site (Fig. 11e) whereas the second event occurred at 10 LST 14 January is composed of the mixture of anthropogenic aerosols and Asian dust aerosol transported from northern China in the dust source region. Although the concentrations of Asian dust aerosol for the second event are small compared with that of anthropogenic aerosols at the surface level (Fig. 11e), the column integrated PM_{10} concentration (not shown) shows that the Asian dust aerosol is predominated for the second event, suggesting the upper level long-range transported Asian dust aerosol being a main cause of the dense haze event at Qingdao.

However, the contribution of Asian dust aerosol to the total

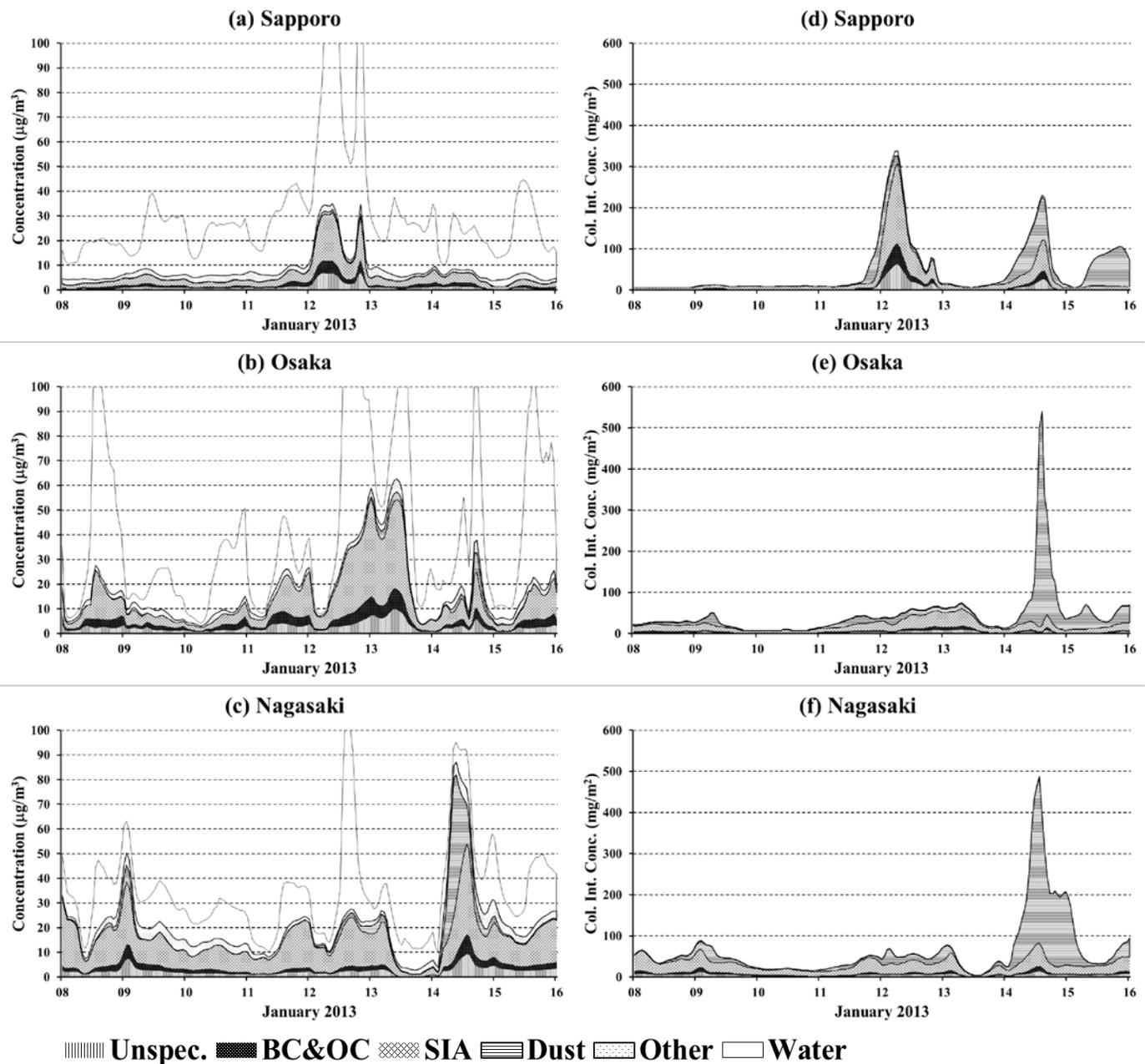


Fig. 12 Time series of model simulated surface PM_{10} concentration ($\mu\text{g}/\text{m}^3$) at (a) Sapporo, (b) Osaka and (c) Nagasaki and the column integrated PM_{10} concentration (mg/m^2) at (d) Sapporo, (e) Osaka and (f) Nagasaki.

PM_{10} concentration is rather small but that of the locally emitted PM_{10} concentration is rather large at Shanghai throughout the whole analysis period (Fig. 11f).

The high aerosol concentration events occurred in China (Figs. 10 and 11) have been transported to Japan with diminished intensity. At the Sapporo site in northern Japan the first anthropogenic aerosol event occurs from 00 LST (= UTC + 9) 12 to 09 LST 13 January (Figs. 12a and d) and then the second mixed aerosol event (anthropogenic and Asian dust aerosols) occurs from 09 LST 14 to 04 LST 15 January with low PM_{10} concentration at the surface (Figs. 12a and d).

A similar feature can be seen at the Osaka site (Figs. 12b and e) located in central Japan and the Nagasaki site (Figs. 12c and f) in Southern Japan with delayed occurrence time by about 12

hours for the first anthropogenic aerosol event. However, the second mixed aerosol event (more dominated by Asian dust aerosol in Figs. 12d, e and f) occurs almost the same time of 00 LST 15 January at all 3 sites, suggesting the upper level long-range transported Asian dust aerosol arriving over Japan almost the same time.

D. Temporal variations of model simulated pollutants concentrations in China and Japan

The anthropogenic aerosol concentration in the Asian dust source region is very low (Figs. 10a, b and c) due to low pollutants emissions (Fig. 2). However, high pollutants emissions in eastern China (Fig. 2) are main sources of the simulated anthropogenic aerosols (Figs 10d, e and f), so that the

behaviors of gaseous pollutants in the high pollutants emission regions are important to understand air quality in these regions.

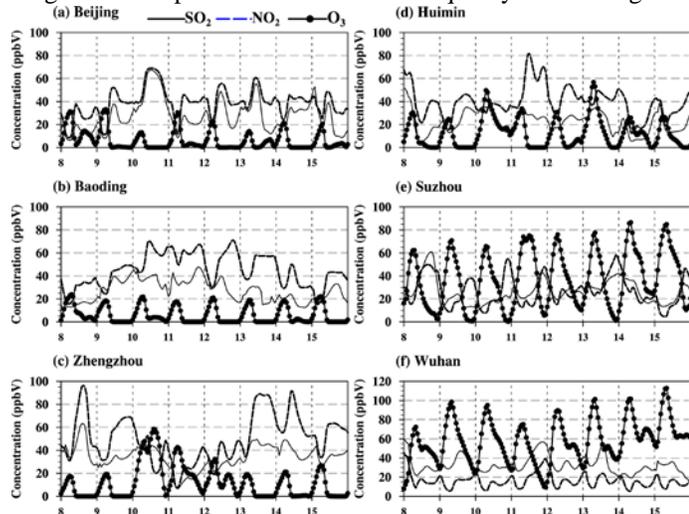


Fig. 13 Time series of model simulated SO_2 (solid line), NO_2 (dashed line), O_3 (solid line with close circle) concentration (ppbV) at (a) Beijing, (b) Baoding, (c) Zhengzhou, (d) Huimin, (e) Suzhou and (f) Wuhan for the period from 8 to 16 January 2013.

Fig. 13 shows time series of simulated concentrations of SO_2 , NO_2 and O_3 at the sites in China. The maximum concentrations of SO_2 and NO_2 exceeding 60 ppb occur at the late night of 10 January after the occurrence of the maximum anthropogenic aerosol concentration at the Beijing site (Fig.10d). However, the O_3 concentration shows a significant diurnal variation with a maximum during mid-day when the NO_2 concentration is minimum whereas near zero O_3 concentration during the early morning and night time when the NO_2 concentration is a maximum of more than 40 ppb (Figs. 13a, b, c and d). This is attributed to the large amount of NO_x emission at these sites. The photolysis of NO_2 results in production of O_3 during the daytime while the O_3 concentration is consumed in the oxidation of NO to produce NO_2 during the nighttime.

It is interesting to note that when the NO_2 concentration

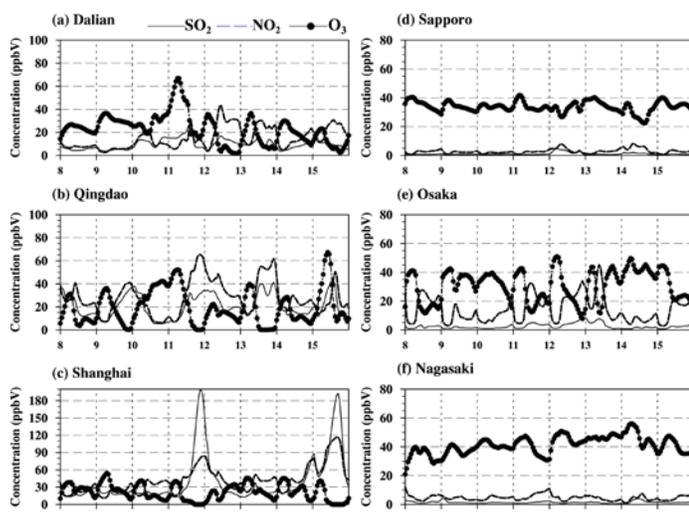


Fig. 14 The same as in Fig. 13 except at (a) Dalian, (b) Qingdao and (c) Shanghai in China, and at (d) Sapporo, (e) Osaka and (f) Nagasaki in Japan.

exceeds 40 ppb the O_3 concentration becomes near zero at the sites of Beijing (Fig. 13a), Baoding (Fig. 13b), Zhengzhou (Fig. 13c) and Huimin (Fig. 13d) where NO_x emission is large (Fig. 2). Therefore, air quality at these sites is largely determined by the anthropogenic aerosols (Fig. 10), NO_2 concentration (Figs 13a, b, c and d) and sometimes high SO_2 concentration whereas O_3 concentration is of no significance.

However, the sites of Suzhou (Fig. 13e) and Wuhan (Fig. 13f) the NO_x emission is relatively low (Fig. 2) so that the NO_2 concentration is relatively low but the O_3 concentration reaches as high as 100 ppb during the daytime causing severe smog events with the high anthropogenic aerosols (Figs. 11b and c).

The pollutants concentrations at the coastal site of Dalian (Fig. 14a) show that SO_2 and NO_2 concentrations are relatively low but the O_3 concentration exceeding 60 ppb occurs during the high anthropogenic aerosol concentration event period from 06 LST 11 to 00 LST 12 January (Fig. 11d). However, those at the sites of Qingdao (Fig. 14b) and Shanghai (Fig. 14c) show high NO_2 and SO_2 concentrations during the high anthropogenic aerosol event periods but the O_3 concentration at both sites is relatively low, suggesting large amount of SO_2 and NO_x emissions (Fig. 2).

The pollutants emissions in Japan are so low (Fig. 2) that concentrations of SO_2 and NO_2 are also low (Figs. 14d, e, and f). The O_3 concentration shows a significant diurnal variation with a mean of 35 ppb at Sapporo (Fig. 14d), 33 ppb at Osaka (Fig. 14e) and with an increasing trend from 23 ppb to 50 ppb at Nagasaki (Fig. 14f) for the analysis period. The long-range transported pollutants cannot be seen in Figs. 10d, e, and f, in contrast to the aerosols that are mainly contributed by the long-range transported ones (Fig. 12).

V. HORIZONTAL DISTRIBUTIONS OF MEAN CONCENTRATIONS FOR THE PERIOD FROM 00 UTC 10 TO 00 UTC 16 JANUARY 2013

A. The horizontal distribution of mean aerosol (PM_{10}) concentration

Fig. 15 shows surface mean total aerosol concentration (PM_{10}) averaged for the period from 00 UTC 10 to 00 UTC 16 January 2013. The aerosols originated in Asia affect all Asian regions including the Asian continent, the northwestern Pacific Ocean, the South China Sea and the Bay of Bengal. The high aerosol concentration exceeding $100 \mu\text{g}/\text{m}^3$ covers over eastern China, the north of the Tibetan Plateau and northern India (Fig. 15). Those high aerosol concentrations over eastern China and northeastern India are mainly contributed by the anthropogenic aerosols (Fig. 16) produced from high pollutants emissions (Fig. 2), and have caused the dense haze events in these regions (Fig. 4). Whereas the high aerosol concentrations over northwestern China to the north of the Tibetan Plateau are mostly contributed by Asian dust aerosol emitted in the Asian dust source region (compare Fig. 15 and Fig. 16).

Fig. 17 shows mean atmospheric aerosol loadings (column integrated aerosol (PM_{10}) concentration) averaged for the period from 00 UTC 10 to 00UTC 16 January 2013. Aerosols produced in the Asia continent affect all over analysis domain

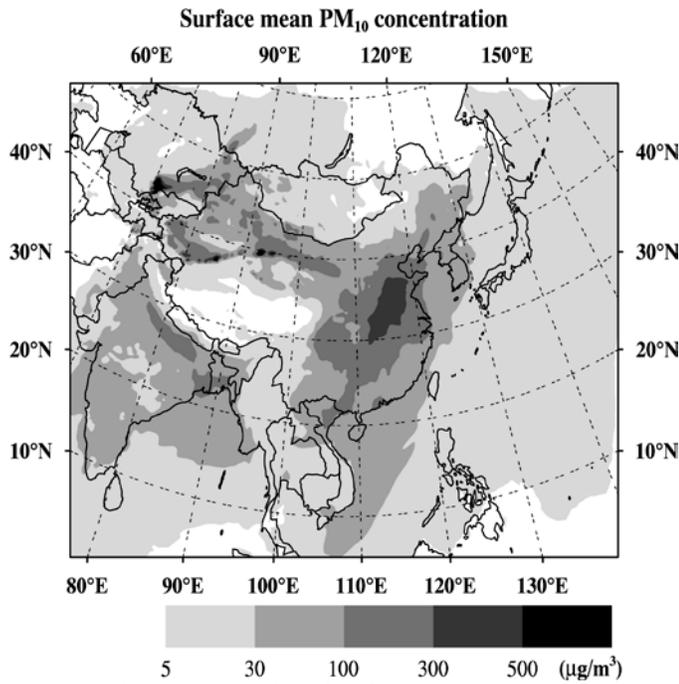


Fig. 15 Horizontal distribution of surface mean PM_{10} concentration ($\mu g/m^3$) averaged for the period from 00 UTC 10 to 00 UTC 16 January 2013.

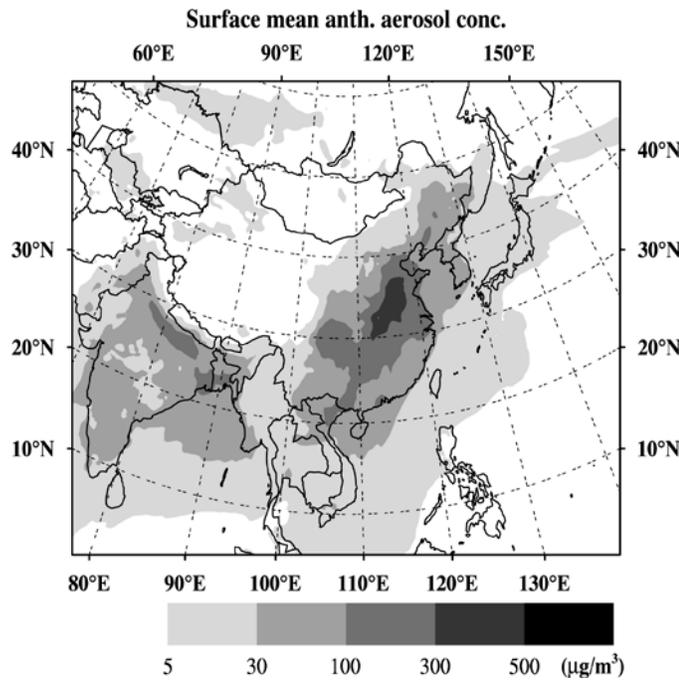


Fig. 16 The same as in Fig. 15 except for the surface mean anthropogenic aerosol concentration ($\mu g/m^3$).

(Fig. 17a) except in the southeastern corner of the model domain over the subtropical high pressure zone.

The high atmospheric aerosol loading zone exceeding 100 mg/m^2 extends southeastward from the Taklamakan desert area of the Asian dust source region to eastern China where it merges with another high aerosol loading zone extending northeastward from the South China Sea through the Korean peninsula to the East Sea of Korea. The high aerosol loading of more than 100 mg/m^2 over northwestern China to the north of the Tibetan

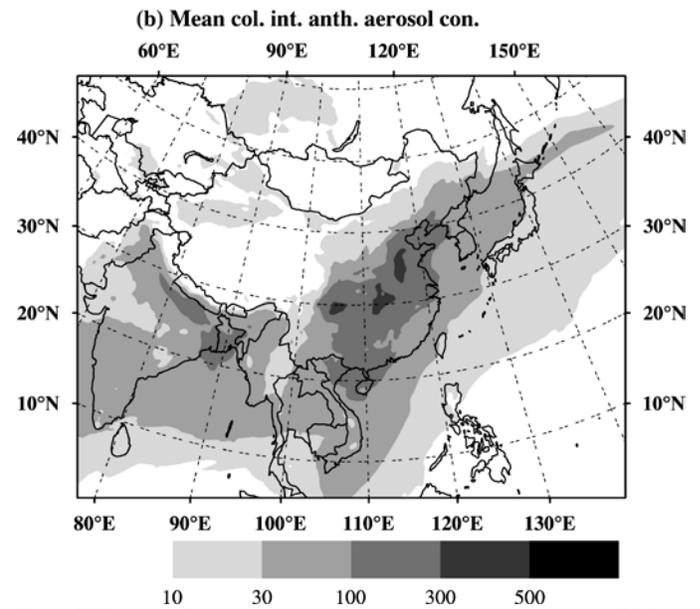
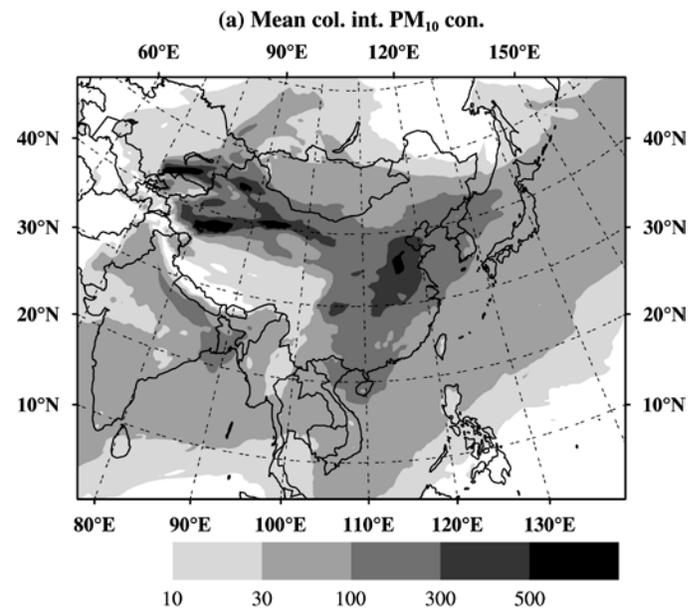


Fig. 17 Horizontal distributions of column integrated mean (a) PM_{10} concentration (mg/m^2) and (b) anthropogenic aerosol concentration (mg/m^2) averaged for the period from 00 UTC 10 to 00 UTC 16 January 2013.

plateau is mainly attributed to Asian dust aerosol while those over eastern China and Northern India are composed of the anthropogenic aerosols (Figs. 17a and b).

It is worthwhile to note that 100 mg/m^2 isoline of the column integrated mean anthropogenic aerosols concentration (Fig. 17b) extends up to the northern part of the Yellow Sea but that of the total aerosol concentration (Fig. 17a) extends further downwind to the East Sea through the Yellow Sea and the Korean peninsula, suggesting a significant amount of Asian dust aerosol being contributed to the total aerosol concentration over the Yellow Sea, the Korean peninsula and the East Sea.

B. Horizontal distributions of surface mean pollutants concentrations averaged for the period from 00 UTC 10 to 00 UTC 16 January 2013

Fig. 18a shows the surface mean SO₂ concentration averaged for the analysis period (00 UTC 10 to 00 UTC 16 January). The horizontal distribution pattern of the SO₂ concentration just follows that of SO₂ emissions (Fig. 2a) and is similar to that of the mean surface anthropogenic aerosol concentration (Fig. 16). The downwind extension of high concentration is quite limited over the source region. Mean surface SO₂ concentration of more than 10 ppb occurs in central eastern China and northern India where the anthropogenic aerosols concentrations are high (Fig. 16).

The horizontal distribution of the surface mean NO₂ concentration (Fig. 18b) also follows that of NO_x emission (Fig. 2b). The mean surface NO₂ concentration exceeding 10 ppb

occurs in the northeastern part of China, along the eastern coastal line of China, the western part of Korea, the northwestern part of India and some big cities in Japan where the NO_x emission is high (Fig. 2b).

The horizontal distribution of the surface mean O₃ concentration (Fig. 18c) is reversed to those of NO₂ concentration (Fig. 18b) and NO_x emission (Fig. 2b). The lowest O₃ concentration occurs over the high NO₂ concentration regions including the northeastern part of China, Shanghai and Seoul in Korea, where NO_x emission is high (Fig. 2b). The high mean surface O₃ concentration exceeding 50 ppb occurs over oceans including the East China Sea, the northwestern Pacific Ocean to the south of Japan, the South China Sea and the Bay of Bengal (Fig. 18c) where the NO_x emission is negligible (Fig. 2b).

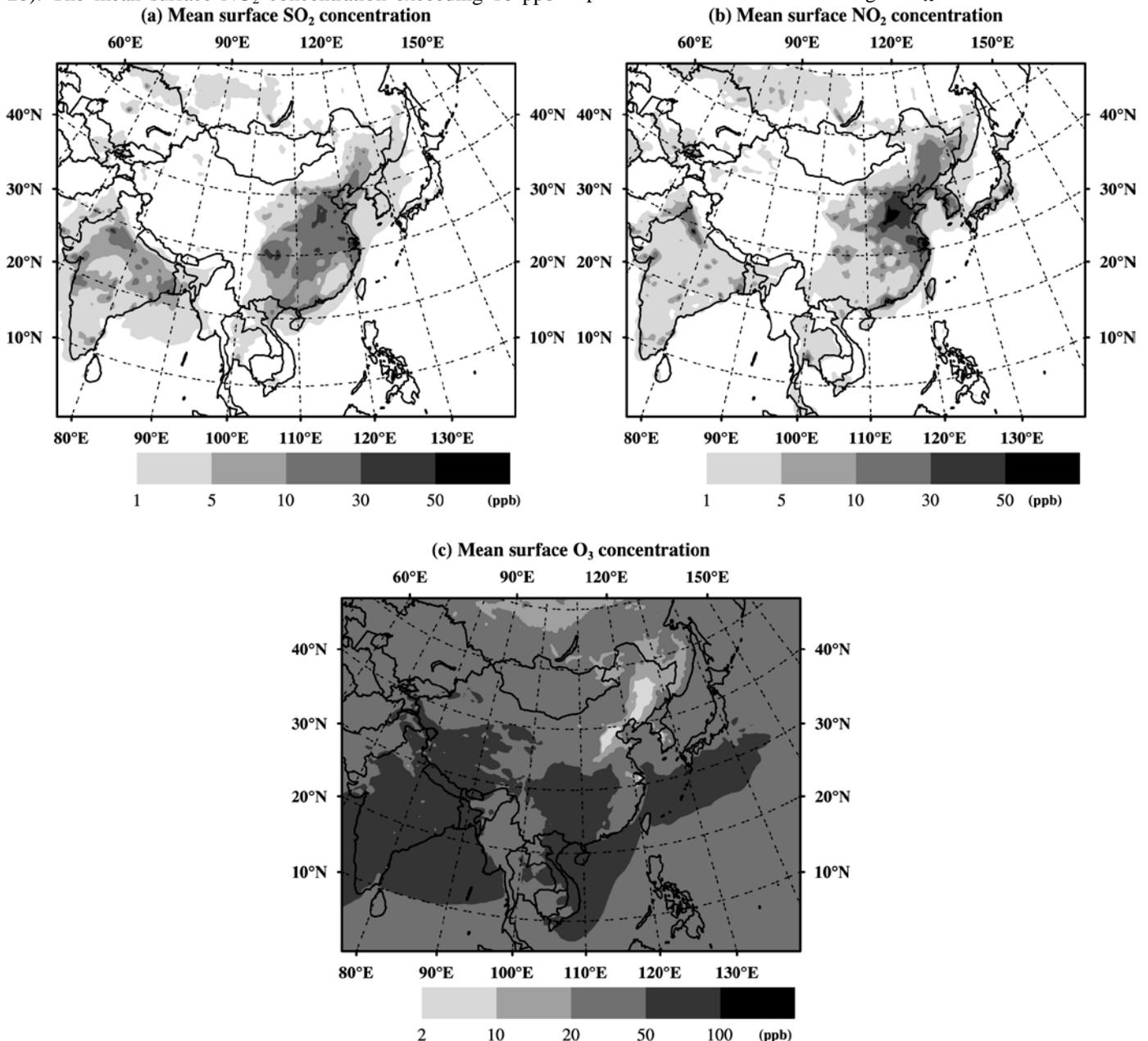


Fig. 18 The same as in Fig. 15 except for the surface mean (a) SO₂, (b) NO₂ and (c) O₃ concentration (ppb)

regions during the daytime are transported over the oceans due to the land breeze during the nighttime and the prevailing westerlies. The transported high O_3 concentrations are conserved over the oceans where there are no NO_x emissions. Consequently, high O_3 concentrations regions extend from the coasts of land regions to the open oceans (Fig. 18c).

It is worthwhile to note that the gas-phase pollutants of SO_2 and NO_2 (Figs. 18a and b) cannot be transported farther downwind direction and their maximum concentrations occurs near their maximum emission regions, whereas the O_3 (Fig. 18c) can be transported further downwind regions sometimes causing a high elevated O_3 concentration over the oceans.

VI. HORIZONTAL DISTRIBUTION OF TOTAL AEROSOL DEPOSITION FOR THE PERIOD FROM 00 UTC 10 TO 00 UTC 16 JANUARY 2013

The deposition of aerosols affects significantly marine and terrestrial ecosystems. Fig. 19a shows the total deposition (wet+dry) of anthropogenic and Asian dust aerosols for the period of the haze event in eastern Asia from 00 UTC 10 to 00 UTC 16 January 2013. The total aerosol deposition of more the 300 kg/km^2 occurs most of Asian dust source region to the north of the Tibetan plateau, Yunnan province, the Beibu gulf, the western coast of Japan, a wide region extending southwestward from the Northwestern Pacific Ocean to Philippines (Fig. 19a).

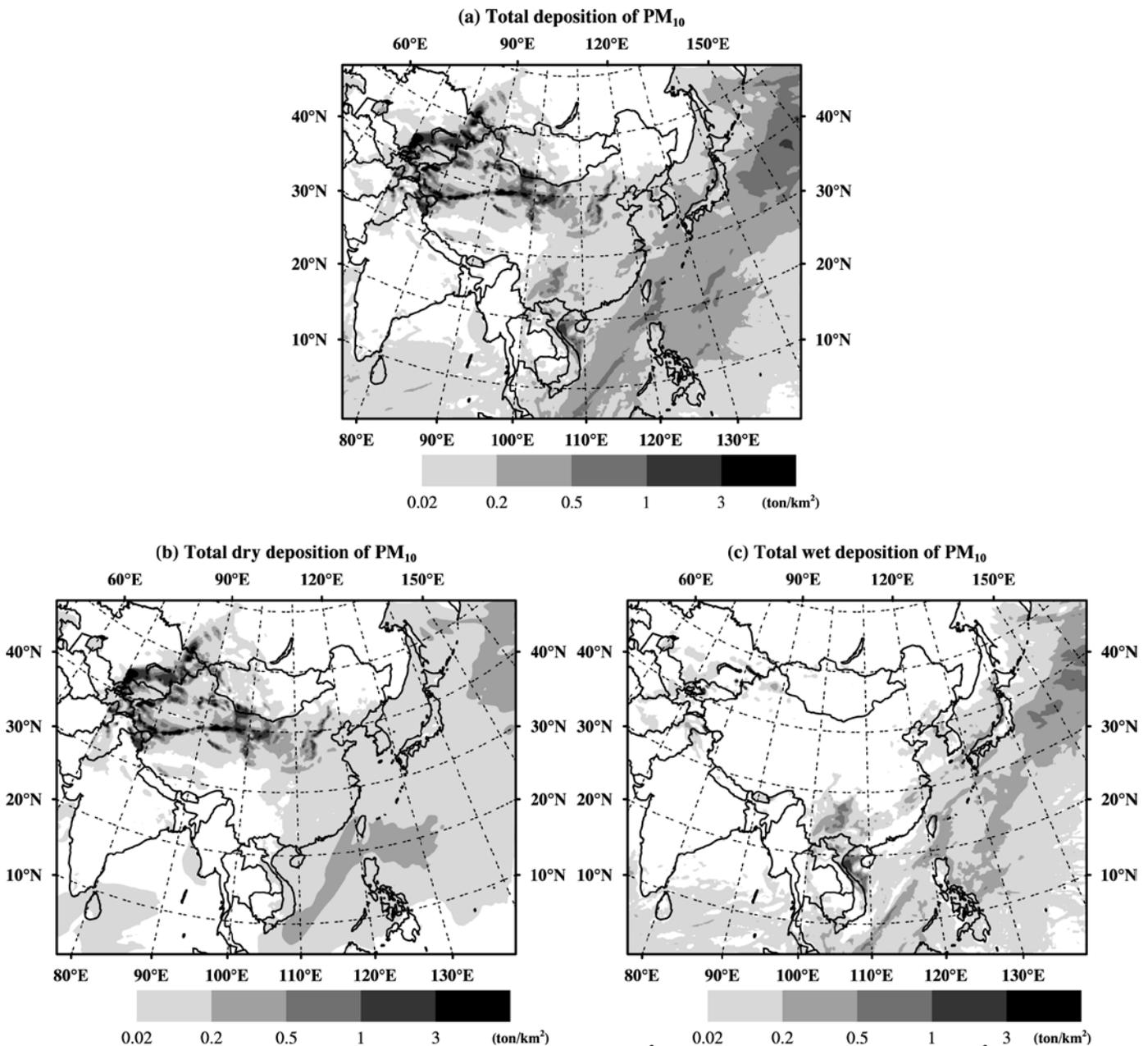


Fig. 19 Horizontal distributions of (a) total deposition (wet+dry) of PM_{10} (t/km^2), (b) total dry deposition of PM_{10} (t/km^2) and (c) total wet deposition of PM_{10} (t/km^2) for the period from 00 UTC 10 to 00 UTC 16 January 2013.

Most of high aerosol deposition to the north of the Tibetan plateau in the Asian dust source region is largely contributed by the dry deposition (Fig. 19b) whereas all other high aerosol deposition regions including Yunnan province, the Beibu gulf, the western coast of Japan and over oceans (the Northwestern Pacific Ocean, the South China Sea) are mainly contributed by wet deposition (Fig. 19c). The large amount of aerosol input more than 1000 kg/km² to the Seas from the Northwestern Pacific Ocean, through the East China Sea to the South China Sea, could affect significantly marine ecology in this region (Fig. 19a).

The present analysis indicates that the high Asian dust emission in northern China and Mongolia and high pollutant emissions in China could produce severe environmental problems including dense haze events that cause severe visibility reduction and adverse impacts on human health through the increased atmospheric aerosol loadings and on terrestrial and marine ecologies through the excess deposition of aerosols.

VII. CONCLUSIONS

The Aerosol Modeling System (AMS) based on the ADAM2 model for the Asian dust aerosol and the CMAQ model for the anthropogenic aerosols has been employed to simulate air quality in Asia in association with dense haze events observed in East Asia for the period from 9 to 16 January 2013. These events caused for the Environmental Protection Bureau in Beijing, China to take emergency measures. The simulated air quality including the gas-phase pollutants (SO₂, NO₂ and O₃) and aerosols (PM₁₀; Asian dust aerosols and anthropogenic aerosols) concentrations has been compared with the monitored gas-phase pollutants and the monitored PM₁₀ concentrations at several sites scattered over South Korea to ensure the usefulness of the model for the further analysis of air quality in Asia.

It is found that the AMS model can simulate quite reasonably both the observed PM₁₀ concentrations and the pollutants concentrations during the period of the dense haze events that have caused severe air pollution problems in China and the downwind region of Korea, and that can identify the contribution of each type of aerosols to these poor air quality events.

It is also found that the dense haze events associated with high aerosol concentration events observed in Korea are caused by two different transport processes; the one is mainly affected by the anthropogenic aerosols originated from the pollutants emitted in eastern China while the other is affected by a mixture of the Asian dust aerosol from northern China in the Asian dust source region and the anthropogenic aerosols from eastern China of the high pollutant emission region. The anthropogenic aerosols during the event period are found to be primarily contributed by the secondary inorganic aerosols, suggesting the importance of the emission of aerosol precursors including SO₂, NO_x and NH₃.

Even though high aerosol concentrations are found to be in the north-western parts of China, the haze events are not found

to be reported in this region. This is due to the fact that the high aerosol concentrations are mainly composed of the Asian dust aerosols with very low relative humidity (water vapor mixing ratio) in the Asian dust source region. However, the severe haze (smog) events occurred in the eastern and southern parts of China including Beijing, Baoding, Huimin, Suzhou, Wuhan and Shanghai are found to be primarily contributed by anthropogenic aerosols originated from pollutants, especially secondary inorganic aerosols. The high pollutants concentrations of SO₂ and NO₂ are also found to be partially contributed to the poor air quality in these regions.

The gas-phase air pollutants of SO₂ and NO₂ are unable to be transported further downwind direction with their maximum concentrations near their maximum source regions. However, aerosols and photochemically produced ozone are found to be transported to the far downwind region. The high aerosol concentrations observed in Korea and Japan for this haze event are primarily contributed by the aerosol produced in China. The ozone produced in the high emission of ozone precursors during the daytime is found to be transported to the far downwind region of the ocean where the NO_x emission is negligible. This sometimes causes high O₃ concentration over the ocean.

The atmospheric aerosol loadings and the total deposition of aerosols in association with these events clearly indicate that not only the air quality in the whole East Asia regions but the marine and terrestrial eco-systems could be significantly affected by these events.

In view of the increasing occurrence trend of such events in East Asia, some reduction measures of pollutants emissions, especially in China are required for keeping sustainable environment in East Asia.

This study mainly pertains to a dense haze event case to understand the impact of the each type of aerosols on this event but further studies for a year-long period are required to assess the impact of pollutants emissions on environment and eco-systems in the Asia region.

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